

Raglan Wastewater Options

Watercare Services Limited Report 44801462/02 April 2021

Raglan Wastewater Options

Prepared for Watercare Services Limited Represented by Mr Stephen Howard

Wainui Stream

CONTENTS

Executive Summary

A previously calibrated Raglan Harbour model has been used to assess a number of potential future discharge options for the Raglan WWTP.

A draft report detailing the model setup, calibration and model results was peer reviewed by Metocean on behalf of the Waikato Regional Council.

That review process identified a number of points of clarification around some of the technical details of the various models used in the study.

As a results of the review process, the draft report was finalised to include a more detailed methodology section, clarification that data from a 1995 current meter deployment was at the existing outfall location (in a similar water depth to the current bathymetry), details of an additional calibration of the hydrodynamic model against observed water levels at Manu Bay, details of an additional validation of currents at a site within the main channel of the harbour (some 300 m from the existing outfall), clarification of how the near-field and farfield models are couple to provide conservative estimates of dilution in the immediate vicinity of the outfall and clarification on the purpose of the wave model.

The predicted level of dilution achieved by the existing outfall for the current day discharge regime is used to benchmark the future options.

These options include a discharge to Wainui Creek (with MBR and UV treatment) and discharges via a proposed new outfall located approximately 100 m to the east of the existing outfall, extending 85 m offshore in a minimum water depth of 2.5 m (compared to a minimum water depth of 0.3 m for the existing outfall).

For the new outfall options, two different levels of treatment were considered - the first being a combination of Pond plus Tertiary Membrane plus UV and the second being a combination of MBR and UV. A full discharge via the new outfall was considered as well as a combination of land disposal and discharge of residual treated wastewater via the new outfall were considered.

The timing of the proposed (and existing) discharges via the outfall have been optimised to maximise the dilution achieved at the outfall sites.

The most optimal discharge window is for a discharge to commence one hour after high water for a period of 4 hours.

Modelling of the dynamics of the treated wastewater plume in the immediate vicinity of the new outfall, show that for the majority of the discharge window the plume would sit in the top 50% of the water column.

The minimum dilution achieved over the existing outfall for the 2025 discharge scenario is 314. This is higher than the minimum level of dilution of 96 achieved for the existing outfall and the non-optimised current discharge regime. This indicates the clear benefit of the optimised discharge timing.

The minimum dilution achieved over the new outfall for the 2055 discharge scenario is 105 – on a par with the level of dilution for the existing outfall and the non-optimised current discharge regime. That is, the improved performance of the new outfall offsets the effects of the increased discharge volume that may occur through to 2055.

There are clear advantages associated with the partial disposal of the treated wastewater to land with reduced (or no) discharges via the new outfall from November through to April.

In addition to considering the level of dilution achieved for the various discharge options, the relative role of catchment and WWTP derived Total Nitrogen have been assessed.

Increases in mean annual Total Nitrogen near the outfall sites increase by less than 0.10 mg/L (compared to background levels of 0.14 mg/L).

For the Wainui Stream option an increase in mean annual Total Nitrogen of 0.12 mg/L is predicted to occur (compared to background levels of 1.01 mg/L).

Data from the calibrated model have been extracted at key sites as input to the Quantitative Microbial Risk Assessment of the future options.

1 Introduction

This report provides details of the use of a calibrated model of Raglan Harbour (DHI, 2019) to assess alternative discharge options for the Raglan wastewater treatment plant. All options are benchmarked against the discharge via the existing outfall located near the entrance to Raglan Harbour.

The options considered include a discharge to the Wainui Stream and discharges via a new outfall located just to the east of the existing outfall. The extended outfall option also considers the discharge of the residual treated wastewater from two land disposal options.

The report provides details of the optimisation of the timing of the outfall discharge (Section [2\)](#page-7-0), modelling of the near-field performance of the new outfall (Section [4\)](#page-23-0), an assessment of the relative roles of the input of nitrogen from the Raglan catchment and the discharge options (Section [7\)](#page-67-0), an overview of the treated wastewater plume dynamics (Section [5\)](#page-29-0) and a summary of the level of dilution achieved at a number of key sites in the Wainui Stream and Raglan Harbour [\(6\)](#page-46-0).

The discharge scenarios and options considered are summarised in [Table 1-1.](#page-7-1)

The current discharge rate for the existing outfall option is based on monitoring data from 2015-2019. Discharge rates for the other options are based on future population projections and the estimated volumes to land for the Public and Private land disposal options (detailed in Section [3\)](#page-18-0).

As for the previous work (DHI, 2019) a combination of near-field modelling and far-field modelling has been used to assess the level of dilution achieved in the immediate vicinity of the discharges and in the wider harbour.

A draft report detailing the model setup, calibration and model results was peer reviewed by Metocean on behalf of the Waikato Regional Council.

That review process identified a number of points of clarification around some of the technical details of the various models used in the study.

As a results of the review process, the draft report was finalised to include a more detailed methodology section, clarification that data from a 1995 current meter deployment was at the existing outfall location (in a similar water depth to the current bathymetry), details of an additional calibration of the hydrodynamic model against observed water levels at Manu Bay, details of an additional validation of currents at a site within the main channel of the harbour (some 300 m from the existing outfall), clarification of how the near-field and farfield models are couple to provide conservative estimates of dilution in the immediate vicinity of the outfall and clarification on the purpose of the wave model.

The near-field modelling has been done using the industry standard CORMIX model (Doneker and Jirka, 2007). The far-field modelling has been carried out using the MIKE3 three-dimensional hydrodynamic and advection-dispersion models which have been coupled to the MIKE21 spectral-wave model to ensure the potential effects of waves on near-shore currents are adequately resolved (as detailed in DHI, 2019).The far-field model was run for the 2018 calendar year, since 2018 is very representative of the long-term distribution of winds, waves, water level variations and freshwater inflows that occur.

Table 1-1 Summary of discharge options considered.

2 Discharge Timing Optimisation

The current consent allows for a discharge to occur half an hour before high water for up to six hours. High water time is taken from the Wharf Tide Gauge (which is the LINZ reference site for Raglan). High tide at the outfall site occurs approximately 35 minutes prior to high water at Raglan Wharf.

Previous work (DHI, 2019) showed that the discharges via the existing outfall often occur prior to high water. When this occurs, the treated wastewater plume is initially transported into Raglan Harbour leading to elevated concentrations just inshore of the outfall.

Earlier modelling showed that maximum predicted concentrations inshore of the outfall are determined by the discharge start time and an analysis of three years of plant discharge data shows that the discharge starting before high water occurs for around 30% of the time.

Even though the public health risk at sites inshore of the outfall are considered to be below the no observable adverse effects level¹, it is recommended that optimising of the start time of the discharge should be carried out to bring about improvements to water quality in the harbour without leading to higher concentrations in areas offshore of the outfall.

To do this, six discharge timings have been considered. All timings refer to local high water at the outfall.

For all options a maximum discharge rate of 3000 m³/day has been assumed, which is the maximum discharge rate being considered for the future options.

Option 1. Current discharge window for 6 hours. Starting half an hour before local high water. Constant discharge rate of 0.064 m³/s.

Option 2. Five hour discharge window. Starting half an hour after local high water. Constant discharge rate of 0.076 m $\frac{3}{s}$.

Option 3. Four hour discharge window. Starting one and a half hours after local high water. Constant discharge rate of 0.083 m³/s.

Option 4. Three hour discharge window. Starting one and a half hours after local high water. Constant discharge rate of $0.119 \text{ m}^3\text{/s}$.

Option 5. Two hour discharge window. Starting two hours after local high water (timed to coincide with peak tidal currents). Constant discharge rate of $0.167 \text{ m}^3/\text{s}$.

Option 6. Four hour discharge window. Starting one hour after local high water. Constant discharge rate of 0.083 m $\frac{3}{s}$.

An example of the timing of the discharge timing options relative to the tide at the outfall site is shown in [Figure 2-1.](#page-10-0)

The previously calibrated harbour model was run for each of the timing options for a 7-day period (starting at a mean tide through to a spring tide). No winds or waves were considered.

A six-hour discharge starting half an hour before high water results in the plume initially being transported inshore of the discharge point. As the tide falls, the plume is transported away from the discharge along Ngarunui Beach [\(Figure 2-2\)](#page-11-0).

A five-hour discharge starting half an hour after high water avoids the elevated concentrations inshore of the outfall. As the tide falls, the plume is transported away from the discharge along Ngarunui Beach and, because of the slightly higher discharge rate, concentrations along northern end of Ngarunui Beach are slightly higher than for the sixhour discharge [\(Figure 2-3\)](#page-12-0)

A four-hour discharge starting one and a half hours after high water avoids the elevated concentrations inshore of the outfall. As the tide falls, the plume is transported away from the discharge along Ngarunui Beach. Because the discharge is happening more towards the peak of the tidal currents, the predicted concentrations along the northern end of Ngarunui Beach are slightly lower than for the six-hour or five-hour discharge options despite the discharge rate being higher [\(Figure 2-4\)](#page-13-0).

The three hour and two hour discharge options result in higher concentrations along northern end of Ngarunui Beach – the effect of the higher discharge rates for these options

¹ NIWA 2019. Human health risk assessment Raglan WWTP. NIWA Client Report 2019297HN prepared for Beca.

is not offset by the discharge occurring when tidal currents are at a maximum [\(Figure 2-5](#page-14-0) and [Figure 2-6\)](#page-15-0).

Finally, the plume dynamics for the four hour discharge commencing at high water plus one hour [\(Figure 2-7\)](#page-16-0) is very similar to those for the four hour discharge commencing one and a half hours after high water [\(Figure 2-4\)](#page-13-0).

Based on this schematic discharge regime of 3000 m³/day, the four hour discharge commencing at local high water plus one hour provides the best overall performance in terms of the predicted maximum concentrations immediately inshore of the outfall [\(Figure](#page-17-0) [2-8\)](#page-17-0) and it does not lead to increases in concentrations offshore of the outfall and along Ngarunui Beach seen with shorter duration discharge options.

Compared to the current consented discharge window timing, this timing option provides more than a five times decrease in the maximum predicted concentration immediately inshore of the outfall and around a two times decrease in the maximum predicted concentration towards the northern end of Ngarunui Beach.

This timing option also avoids the discharge occurring towards local low water when a combination of shallow water depth and reduced tidal currents result in relatively low levels of dilution occurring in the immediate vicinity of the outfall and the highest visual impact from the discharge.

Figure 2-1. Example timings for the six timing options considered.

Figure 2-2. Six-hour discharge option commencing half an hour prior to high water (the current consent discharge window). Predicted concentration at the start of the discharge (half an hour prior to local high water), just prior to peak-ebb tide currents and just after peak ebb tide currents.

Figure 2-3. Five-hour discharge option commencing half an hour after high water. Predicted concentration at the start of the discharge (local high water plus half an hour), just prior to peak-ebb tide currents and just after peak ebb tide currents.

Figure 2-4. Four-hour discharge option commencing one and a half hours after high water. Predicted concentration at the start of the discharge (local high water plus one and a half hours), just prior to peak-ebb tide currents and just after peak ebb tide currents.

Figure 2-5. Three-hour discharge option commencing one and a half hours after high water. Predicted concentration at the start of the discharge (local high water plus one and a half hours), just prior to peak-ebb tide currents and just after peak ebb tide currents.

Figure 2-6. Two-hour discharge option commencing two hours after high water. Predicted concentration at the start of the discharge (local high water plus two hours - just prior to peak-ebb tide currents) and just after peak ebb tide currents.

Figure 2-7. Four-hour discharge starting at local high water plus one hour. Predicted concentration at the start of the discharge (high water plus one hour), just prior to peak-ebb tide currents and just after peak ebb tide currents.

Figure 2-8. Predicted maximum concentrations just inshore of the outfall and at the very northern end of Ngarunui Beach for the timing options considered. Timing options are sorted by lowest predicted maximum concentration at the inshore site.

3 Discharge Scenarios

The following provides the assumptions used to derive the discharge rates for the four discharge options being considered for the Raglan Wastewater Treatment Plant.

These options are:

- A discharge via the exiting outfall for 2020 flow rates;
- A discharge via the new outfall for estimated flow rates in 2025 and 2055;
- A continuous discharge to the Wainui Stream for estimated flow rates in 2025 and 2055;
- A Public Land disposal option with partial discharge to a new outfall for estimated flow rates in 2025 and 2055; and
- A Private Land disposal option with partial discharge to a new outfall for estimated flow rates in 2025 and 2055

All options consider the predicted 2025 and 2055 Average Dry Weather discharges of 1372 and 1957 m³/day respectively.

The average daily flow for the period from 2015 through to 2019 is 1025 m $3/$ day with the monthly variation as shown in [Table 3-1.](#page-19-0) These volumes are used for the Existing scenario [\(Table 1-1\)](#page-7-1) via the existing outfall.

This variability is used to define the mean monthly flows for the 2025 and 2055 discharge scenarios [\(Table 3-2\)](#page-19-1).

These discharge volumes are used Options M1, M2 (via a new Outfall - [Table 1-1\)](#page-7-1) and F1 (via Wainui Stream - [Table 1-1\)](#page-7-1). For the outfall option the discharge occurs over four hours commencing one-hour after local high water.

PDP supplied estimates of the volumes that could be disposed of via both the Public and Private Land disposal options [\(Table 3-3](#page-20-0) and [Table 3-4\)](#page-20-1). For these options, the residual volume would be discharged via an outfall (with the same 4 hour timing as for Options M1 and M2).

The volume to the new outfall for Options L1 and L4 (Public Land Disposal - [Table 1-1\)](#page-7-1) and L4 (Private Land Disposal [-Table 1-1\)](#page-7-1) and are shown in [Table 3-5](#page-21-0) and [Table 3-6](#page-21-1) respectively.

For the Public Land disposal options there is always some discharge to the outfall while for the Private Land disposal option there is no discharge to marine for 7 months (in 2025) and 5 months (in 2055).

An example of the timing of the discharges relative to the state of tide are shown in [Figure 3-1.](#page-22-0)

Table 3-1. Mean daily discharge from the Raglan Wastewater Treatment Plant (2015-2019). These volumes are used for the discharge to the **Existing Outfall**.

Table 3-2. Assumed distribution of monthly mean daily discharge for the 2025 and 2055 discharge scenarios based on the current monthly distribution of discharges [\(Table 3-1\)](#page-19-0). These volumes are used for the discharge scenarios to the **New Outfall** and **Wainui Stream**.

Table 3-3. Assumed potential volumes irrigated to **Public Land** (with 50m Property Buffered Areas and Public Event Spaces Removed). Note that the volumes are all less than the assumed mean monthly discharge from the plant [\(Table 3-2\)](#page-19-1) so that residual volumes are discharged via the outfall for all months.

Table 3-4. Assumed potential volumes irrigated to **Private Land**. Note that these volumes are often more than the assumed mean monthly discharge from the plant [\(Table 3-2\)](#page-19-1) so that for some months there will is no marine discharge component.

Table 3-5. Assumed distribution of monthly mean daily discharge to **New Marine Outfall** for the 2025 and 2055 discharge scenarios with **Public Land Disposal.**

Table 3-6. Assumed distribution of monthly mean daily discharge to **New Marine Outfall** for the 2025 and 2055 discharge scenarios with **Private Land Disposal**.

Figure 3-1. Example timing for the new outfall option for the 2055 discharge rate.

4 Near Field Modelling

The new outfall location [\(Figure 4-1\)](#page-24-0) is approximately 100 m to the east of the existing outfall, 85 m offshore and would sit in a minimum water depth of 2.5 m (compared to a minimum water depth of 0.3 m for the existing outfall).

For the new outfall near-field modelling has been done using the industry standard CORMIX model (Doneker and Jirka, 2007). This model considers the configuration of the outfall structure, the discharge flow rate, the discharge characteristics, the bathymetry of the point of discharge and the range of currents and water depths that can occur over an outfall. Outputs from the CORMIX model are used to quantify the behaviour of the treated wastewater discharge plume within the first few hundred metres of the discharge point and to determine when the treated wastewater plume rapidly becomes fully mixed in the water.

Previous near-field modelling of the existing outfall (DHI, 2019) indicated that the lowest level of dilution occurs near high water when ambient currents are the lowest.

Under such conditions, the plume from the existing outfall occupies the top 10% of the water column and the 10-15 fold dilution is achieved at the edge of the near-field (i.e. 10-15 m from the existing outfall). At other phases of the tide, much higher levels of dilutions are achieved, and the plume becomes fully mixed through the water column within 225 metres of the outfall.

By avoiding a discharge near high-water the minimum dilution achieved over the existing outfall site will be much higher than previously modelled.

For this assessment the key outcome of the near-field modelling is to ensure that the plume dynamics within the near-field zone of the new outfall are adequately and conservatively represented in the far-field model. To do this a number of worst-case combinations of low ambient current conditions and minimal associated water depths over the outfall have been assessed using the CORMIX model.

For the new outfall location, the distribution of water depth and current speed over the outfall for the first three hours of the discharge are shown in Figures [4-2](#page-25-0) through to [4-5.](#page-26-0) The plots also show the schematic CORMIX scenarios modelled (as summarised in [Table 4-1\)](#page-27-0).

The CORMIX schematic conditions are modelled for the maximum discharge rate being considered $(2715 \text{ m}^3/\text{day} - \text{Table 3-2})$ and a single port with duck-bill valve.

Because a duck-bill valve is being fitted to the new outfall, the jet velocity will be relatively constant over the range of discharges being considered so that the jet momentum term (which is a key process in defining the near-field mixing) will be similar for the lower discharge rates being considered.

The CORMIX model results [\(Table 4-2\)](#page-28-0) show that, with the exception of the extreme low current scenario of 0.1 m/s (which only occurs ~1% of the time - [Figure 4-2\)](#page-25-1), the dilution achieved over the outfall is 2-3 times higher than the minimum dilution achieved over the existing outfall.

The CORMIX model results also show that the treated wastewater discharge always occupies more than the top 20% of the water column and that the size of the near field region is relatively small (5-10 m) due to the strong tidal currents that occur during the discharge window.

Furthermore, the CORMIX modelling shows that, with the exception of the extreme low current scenario of 0.1 m/s the plume becomes fully vertically mixed within less than 275 m of the new outfall.

As for the earlier work (DHI, 2019), the treated wastewater discharge is conservatively added to just the top 20% of the water column of the far-field hydrodynamic model (i.e. the top layer of the MIKE3 model). This approach is conservative because for the majority of the time the discharge

is occurring the plume will occupy more than the top 20% of the water column in the immediate vicinity of the outfalls and therefore concentrations near the surface will be less than those predicted by the far-field model (i.e. actual dilutions achieved over the new outfall are likely to be higher).

Beyond ~300m of the outfalls the plume is predicted to be fully mixed so any assumptions about the near-field behaviour incorporated into the far-field model do not affect the far-field model results.

Appendix A provides plots of predicted dilution versus distance from the new outfall for the schematic conditions modelled.

Figure 4-1. Existing and proposed new outfall location.

Figure 4-2. Distribution of predicted water depth and current speed over the new outfall during the first hour of the discharge window. The red symbols show the schematic CORMIX scenarios modelled.

Figure 4-3. Distribution of predicted water depth and current speed over the new outfall during the second hour of the discharge window. The red symbols show the schematic CORMIX scenarios modelled.

Figure 4-4. Distribution of predicted water depth and current speed over the new outfall during the third hour of the discharge window. The red symbols show the schematic CORMIX scenarios modelled.

Figure 4-5. Distribution of predicted water depth and current speed over the new outfall during the fourth hour of the discharge window. The red symbols show the schematic CORMIX scenarios modelled.

Table 4-1 Summary of CORMIX scenarios modelled.

Table 4-2 CORMIX results for the schematic scenarios modelled.

5 Plume Dynamics

Figures [5-1](#page-30-0) through to [Figure 5-16](#page-45-0) show the predicted $95th$ percentile plots for each of the scenarios considered.

Percentile plots for the period from January-March and July-September are provided for each scenario so that the influence of the reduced volumes to the outfall for the land disposal options can be visualised against the full discharge to marine options.

For the outfall options, the plots show the clear distinction between the January-March dilutions (when discharge volumes are lower) and those in July-September. The area where dilutions of less than 2000 are predicted to occur extends more offshore during the July-September period compared to in January-March. There is also an overall reduction in the level of dilution achieved in July-September period compared to in January-March.

The reduced volumes to the outfall with the Land Disposal options results in an overall increase in dilution.

For the Wainui Stream options, the zone where dilutions of less than 100 are achieved is very similar for the January-March period and the July-September period (extending slightly more into the Opotoru Arm in July-September).

The area where a dilution of less than 100 is achieved increases slightly between the 2025 scenario and the 2055 scenario from ~9.6 Ha to ~12.0 Ha.

In the wider harbour, as for the outfall options, there is an overall increase in the level of dilution achieved in July-September period compared to in January-March.

Figure 5-1. Predicted 95th percentile dilution for the January-March period for Existing Scenario (Existing Outfall, Current Discharge rate).

Figure 5-2. Predicted 95th percentile dilution for the July-September period for Existing Scenario (Existing Outfall, Current Discharge rate).

Plume Dynamics

Figure 5-3. Predicted 95th percentile dilution for the January-March period for Scenario M1 (New Outfall, 2025 Discharge rate).

Figure 5-4. Predicted 95th percentile dilution for the July-September period for Scenario M1 (New Outfall, 2025 Discharge rate).

Plume Dynamics

Figure 5-5. Predicted 95th percentile dilution for the January-March period form M1 (New Outfall, 2055 Discharge rate).

Figure 5-6. Predicted 95th percentile dilution for the July-September period form M1 (New Outfall, 2055 Discharge rate).
Plume Dynamics

Figure 5-7. Predicted 95th percentile dilution for the January-March period for Scenario L1 (Public Land disposal plus New Outfall, 2025 Discharge rate).

Figure 5-8. Predicted 95th percentile dilution for the July-September period form L1 (Public Land Disposal plus New Outfall, 2025 Discharge rate).

Plume Dynamics

Figure 5-9. Predicted 95th percentile dilution for the January-March period for Scenario L3 (Public Land disposal plus New Outfall, 2055 Discharge rate).

Figure 5-10. Predicted 95th percentile dilution for the July-September period form L3 (Public Land Disposal plus New Outfall, 2025 Discharge rate).

Plume Dynamics

Figure 5-11. Predicted 95th percentile dilution for the July-September period form L3 (Private Land Disposal plus New Outfall, 2025 Discharge rate). Note there is no discharge in January-March for this discharge options.

Figure 5-12. Predicted 95th percentile dilution for the July-September period form L3 (Private Land Disposal plus New Outfall, 2055 Discharge rate). Note there is no discharge in January-March for this discharge options.

Plume Dynamics

Figure 5-13. Predicted 95th percentile dilution for the January-March period for Scenario F1 (Wainui Stream, 2025 Discharge rate).

Figure 5-14. Predicted 95th percentile dilution for the July-September period for Scenario F1 (Wainui Stream, 2025 Discharge rate).

Plume Dynamics

Figure 5-15. Predicted 95th percentile dilution for the January-March period for Scenario F1 (Wainui Stream, 2055 Discharge rate).

Figure 5-16. Predicted 95th percentile dilution for the July-September period for Scenario F1 (Wainui Stream, 2055 Discharge rate).

6 Dilutions at QMRA Sites

The following section provides a summary of the dilutions achieved at each of the QMRA sites [\(Figure 6-1\)](#page-47-0) for each of the discharge options and discharges considered [\(Table 1-1\)](#page-7-0).

The minimum dilution achieved over the existing outfall for the 2025 discharge scenario is 314 This is higher than the minimum level of dilution of 96 achieved for the existing outfall and the non-optimised current discharge regime (DHI, 2019) giving a clear indication of the benefit of the optimised discharge timing.

The minimum dilution achieved over the new outfall for the 2055 discharge scenario is 105 – on a par with the level of dilution for the existing outfall and the non-optimised current discharge regime. That is the improved performance of the new outfall (i.e. optimised timing, duck-bill valve and increased water depth) offsets the effects of the increased discharge volume through to 2055.

Taking into account the conservative nature of the assumption around the schematisation of the of the plume in the far-field model (i.e. it is only ever in the top 20% of the water column) actual minimum dilutions over the new outfall could be a factor of 2 times more than has been modelled.

Further improvements to water quality will also be achieved through the proposed higher level of treatment for the discharges from the new outfall [\(Table 1-1\)](#page-7-0) but the overall risk of the future discharges will be assessed as part of the QMRA process.

At the nearest QMRA sites to the outfall an order of magnitude increase in the 99.9th percentile dilution is predicted to occur compared to the level of dilution achieved directly over the outfalls with significant increases in dilution at other QMRA sites (i.e. > 10,000).

For the Wainui Stream option, the 99.9th percentile dilution within the Wainui Stream and Opotoru Arm of the harbour range from 32 to 172 for the 2025 discharge scenario and decrease to between 23 and 120 under the 2055 discharge scenario. The minimum level of dilution achieved reflects the relative flow off the Wainui catchment compared to the treated wastewater discharge volumes being considered.

Within the main body of the harbour, the 99.9th percentile dilutions range from 300 (at Site S14) through to around 15,000 at Site S4 for the 2025 discharge and these decrease to 200 and 11,000 under the 2055 discharge.

Figure 6-1. Sites where model data is extracted for the QMRA. S denotes a Shellfish site and R denotes a recreational site.

Table 6-1 QMRA sites. Confirm with Chris R vs S numbering

Table 6-2 Summary of dilutions for the Existing Scenario (Existing outfall and current discharge volume) within the wider harbour.

99.5 | 759659 | 204444 | 76622 | 38802 | 36719 | 35797 | 35728 | 36107 | 35705 | 35730

99.9 | 432894 | 169962 | 60002 | 32011 | 31618 | 31318 | 31350 | 31323 | 31457 | 31694

Table 6-3. Summary of dilutions for the Existing Scenario (Existing outfall and current discharge volume) within Wainui Stream and the Opotoru Arm.

Table 6-4 Summary of dilutions for Scenario M1 (New outfall and 2025 discharge volume) within the wider harbour.

 \equiv

۰.

T

--

Table 6-6 Summary of dilutions for Scenario M1 (New Outfall and 2055 discharge volume) within the wider harbour.

Table 6-8 Summary of dilutions for Scenario L1 (Outfall in combination with Public Land disposal and 2025 discharge volume) within the wider harbour.

99.9 | 843534 | 206537 | 65300 | 27818 | 27899 | 27222 | 27557 | 27409 | 27750 | 27913

Table 6-9. Summary of dilutions for Scenario L1 (Outfall in combination with Public Land disposal and 2025 discharge volume) within Wainui Stream and the Opotoru Arm.

Table 6-10 Summary of dilutions for Scenario L1 (Outfall in combination with Public Land disposal and 2055 discharge volume) within the wider harbour.

Table 6-11. Summary of dilutions for Scenario L1 (Outfall in combination with Public Land disposal and 2055 discharge volume) within Wainui Stream and the Opotoru Arm.

i.

Table 6-13. Summary of dilutions for Scenario L3 (Outfall in combination with Private Land disposal and 2025 discharge volume) within Wainui Stream and the Opotoru Arm.

Table 6-14 Summary of dilutions for Scenario L3 (Outfall in combination with Private Land disposal and 2055 discharge volume) within the wider harbour.

Table 6-16 Summary of dilutions for Scenario F1 (Wainui Stream option and 2025 discharge volume) within the wider harbour.

Site Description

Wainui Stream (Recreational)

Marae (Shellfish)

Airstrip (Recreational)

(Recreational/Shellfish)

Wainui/Opotoru (Recreation/Shellfish)

Domain Boat Ramp (Recreation/Shellfish)

Domain Boat Ramp (Recrea **Site ID R16 R17/S15 R18 R19 R20/S17 R9/S13 R21 R22 R23 R24 Percentile Dilution 90.0** $\begin{array}{|c|c|c|c|c|c|c|c|} \hline \end{array}$ 40 $\begin{array}{|c|c|c|c|c|c|c|c|} \hline \end{array}$ 43.0 $\begin{array}{|c|c|c|c|c|c|} \hline \end{array}$ 40 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 45 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 56 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 425 $\begin{array}{|c|c|c$ **99.0** $\begin{array}{|c|c|c|c|c|c|} \hline \end{array}$ 34 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 39 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 250 $\begin{array}{|c|c|c|c|} \hline \end{array}$ 156 $\begin{array}{|c|c|c|c|} \hline \end{array}$ 152 $\begin{array}{|c|c|c|c|} \hline \end{array}$ 213 **99.5** $\begin{array}{|c|c|c|c|c|c|} \hline \end{array}$ 33 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 34 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 36 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 223 $\begin{array}{|c|c|c|c|} \hline \end{array}$ 95 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ 141 $\begin{array}{|c|c|c|c|c|} \hline \end{array}$ **99.9** $\begin{array}{|c|c|c|c|c|c|c|c|} \hline \textbf{32} & \textbf{32} & \textbf{33} & \textbf{34} & \textbf{35} & \textbf{179} & \textbf{80} & \textbf{126} & \textbf{122} & \textbf{172} \ \hline \end{array}$

Table 6-17. Summary of dilutions for Scenario F1 (Wainui Stream option and 2025 discharge volume) within Wainui Stream and the Opotoru Arm.

Table 6-18 Summary of dilutions for Scenario F1 (Wainui Stream option and 2055 discharge volume) within the wider harbour.

Table 6-19. Summary of dilutions for Scenario F1 (Wainui Stream option and 2055 discharge volume) within Wainui Stream and the Opotoru Arm.

7 Assessment of Water Column Total Nitrogen

This section of the report provides details of the methodology used and assumptions made to assess the relative roles of catchment derived Total Nitrogen (TN) and the loads from the proposed discharged options.

Catchment loads have been derived by mapping different land uses within the sub-catchments of Raglan Harbour [\(Figure 7-1\)](#page-68-0) and applying minimum and maximum TN yields [\(Table 7-1\)](#page-69-0) as tabulated in WRC (2018). The individual minimum and maximum TN loads for each of the catchments are shown in [Table 7-2.](#page-70-0) Data from Zeldis et al. (2017) estimate a TN river load of 420 tonnes/yr. The average of the minimum and maximum loads in [Table 7-2](#page-70-0) have been scaled to achieve this mean annual catchment derived TN load of 420 tonnes/yr.

Based on the mean flow for each of the catchment sources in the Harbour model a mean concentration was initially set for each of the catchment sources so that a total TN load of 420 tonnes/yr was delivered during the 2018 model simulation. A number of iterations of the harbour model were then run to achieve a good match between the predicted mean annual concentrations from the model and the observed mean TN concentrations at the Waikato Regional Council monitoring sites [\(Figure 7-2\)](#page-71-0). This was achieved by applying an attenuation factor of 85% for all the catchment loads except the Opotoru and Te Terata catchments which had attenuation factors of 43% and 64% applied respectively. An attenuation factor of 85% is within the broad scale range of calibrated attenuation factors applied for TN within the Waikato and Waipa River basin catchments (Semadeni-Davies et al., 2016). The calibration processes indicates that the predicted TN loads for the Opotoru and Te Terata catchments using the above methodology are likely to be too high.

No decay was applied to the TN within the marine receiving environment. This approach is justified because of the strong correlation between observed salinity and observed TN at the Waikato Regional Council monitoring sites. Salinity acts as a conservative tracer and can be used to define the level of dilution achieved for other conservative tracers at a particular site in the harbour. The calibration of predicted mean annual TN concentrations versus the observed mean values is given in [Figure 7-4.](#page-73-0) The model tends to under predict the concentrations at the historic monitoring sites (except for the Mid Harbour site) but overall provides good estimates of the observed at the current monitoring sites (i.e. Slope of regression =1.10 and $r^2 = 0.95$).

The calibrated model loads for each of the catchments is shown in [Table 7-3](#page-72-0) with the total load of 324 tonnes/yr entering the system.

The spatial map of the predicted TN due to the attenuated TN catchment load is shown in [Figure](#page-73-1) [7-5.](#page-73-1)

The calibrated TN model was then rerun with the inputs for the existing outfall (2020 discharge rate) and the 2055 discharge scenarios for the new outfall and the Wainui Stream. The assumed median TN concentrations and annual loads for these scenarios are shown in [Table 7-4.](#page-74-0)

Spatial plots of the predicted mean annual TN footprint and percentage increase in mean annual TN (i.e. percentage increase above the background levels from the catchment derived sources) are shown in Figures [7-6](#page-75-0) through to [7-9.](#page-78-0) Note that the concentration plot colour banding is 1) a semi-log scale and 2) at a different scale to the catchment derived plot.

For the outfall options there is a very small offshore zone where there are increases in TN of the order of 0.01 mg/L for the 2020 Existing and 2055, pond plus tertiary membrane plus UV level of treatment and less than this for the MBR plus UV level of treatment. Catchment derived mean annual TN directly over the outfall is around 0.14 mg/L so these increases equate to 9%, 11% and 4% in TN directly over the discharge location.

For the Wainui Stream option the maximum increase in TN (directly at the point of discharge) is of the order of 0.12 mg/L. The mean annual TN at the Wainui catchment outlet is 1.01 mg/L so the increase due to the 2055 discharge equates to a 12% increase in TN.

Figure 7-1. Raglan Harbour catchments.

Table 7-1 Summary of land use types within each of the Raglan catchments and estimated Total Nitrogen loads (WRC, 2018).

Catchment	Area (Ha)	Total Nitrogen (tonne/yr) Minimum Estimate	Total Nitrogen (tonne/yr) Mean Estimate	Total Nitrogen (tonne/yr) Maximum Estimate	Best Estimate of Mean Total Nitrogen $(tonne/yr)^2$
Waitetuna	16866.2	89.1	221.1	335.0	123.1
Waingaro	12498.1	71.9	176.7	269.0	98.4
Okete Bay	4193.0	35.7	84.5	122.6	47.0
Opotoru	3699.2	34.2	80.7	114.9	44.9
Kerikeri	2128.2	17.0	40.6	59.8	22.6
Ohautira	5109.4	14.4	39.5	64.1	22.0
Te Kotuku	979.7	10.3	24.2	34.2	13.5
Wainui	1298.5	8.8	21.3	30.8	11.8
Poganui	978.0	8.4	19.9	28.5	11.1
Okete Falls	585.7	5.1	12.1	17.5	6.7
Hauroto Bay	419.2	4.5	10.4	14.7	5.8
Te Tarata	426.8	4.1	9.7	13.7	5.4
Tawatahi	287.5	3.0	7.1	10.0	3.9
Total	49469.5	306.7	754.2	1114.9	420.0

Table 7-2 Catchment areas and estimated Total Nitrogen loads.

² The mean load estiamtes scaled to provide a total of 420 tonne/yr as reported in Zelids et al. 2017.

Figure 7-2. Waikato Regional Council monitoring sites.

Figure 7-3. Observed Salinity and Total Nitrogen at all of the Waikato Regional Council monitoring sites [\(Figure 7-2\)](#page-71-0).

Table 7-3 Total Nitrogen load modelled and attenuation factor based on a predicted overall load of 420 tonnes/yr.

Figure 7-4. Predicted mean Total Nitrogen at each of the Waikato Regional Council sites [\(Figure 7-2\)](#page-71-0) versus the mean of all observed data.

Figure 7-5. Mean annual Total Nitrogen (mg/l) for all Raglan harbour catchment sources.

Table 7-4 Assumed median Total Nitrogen concentrations (mg/l) for the options considered.

Figure 7-6. Mean annual Total Nitrogen footprint for the existing Outfall (2020) discharge scenario with Pond and UV treatment (top panel) and increase in Total Nitrogen (as a percentage of the catchment derive mean Total Nitrogen - Figure 5 5).

Figure 7-7. Mean annual Total Nitrogen footprint for the new Outfall (2055) discharge scenario with Pond, Tertiary and UV treatment (top panel) and increase in Total Nitrogen (as a percentage of the catchment derive mean Total Nitrogen - [Figure 7-5\)](#page-73-0).

Figure 7-8. Mean annual Total Nitrogen footprint for the new Outfall (2055) discharge scenario with MBR and UV treatment (top panel) and increase in Total Nitrogen (as a percentage of the catchment derive mean Total Nitrogen - [Figure 7-5\)](#page-73-0).

Figure 7-9. Mean annual Total Nitrogen footprint for the Wainui Stream (2055) discharge scenario with MBR and UV treatment (top panel) and increase in Total Nitrogen (as a percentage of the catchment derive mean Total Nitrogen - [Figure 7-5\)](#page-73-0).

References

Doneker, RL and Jirka GH 2007. CORMIX User Manual: A Hydrodynamic Mixing Zone Model and Decision Support System for Pollutant Discharges into Surface Waters. U.S. Environmental Protection Agency.

DHI 2019. Raglan Wastewater Treatment Plant Discharge Assessment. DHI report 44801398 prepared for Beca.

Semadeni-Davies, A., Sandy Elliott, S. and Yalden, S. 2016. Modelling nutrient loads in the Waikato and Waipa River Catchments. Development of catchment-scale models. NIWA Client Report HAM2015-089 Prepared for Waikato Regional Council.

WRC 2018. Waikato stormwater management guideline. WRC Technical Report 2018/01.

Zeldis, J., Plew, D., Whitehead, A., Madarasz-Smith, A., Oliver, M., Stevens, L., Robertson, B., Burge, O., Dudley, B. 2017. The New Zealand Estuary Trophic Index (ETI) Tools: Web Tool 1 - Determining Eutrophication Susceptibility using Physical and Nutrient Load Data. Ministry of Business, Innovation and Employment Envirolink Tools: C01X1420.

A Appendix A – CORMIX results

Figure A-1. Dilution versus downstream distance for Case A [\(Table 4-1\)](#page-27-0). Dilution versus downstream distance for Case A [\(Table 4-1\)](#page-27-0).

Figure A-2. Dilution versus downstream distance for Case A [\(Table 4-1\)](#page-27-0). Dilution versus downstream distance for Case B [\(Table 4-1\)](#page-27-0).

Figure A-3. Dilution versus downstream distance for Case C [\(Table 4-1\)](#page-27-0).

Figure A-4. Dilution versus downstream distance for Case D [\(Table 4-1\)](#page-27-0).

Figure A-5. Dilution versus downstream distance for Case E [\(Table 4-1\)](#page-27-0).

Figure A-6. Dilution versus downstream distance for Case F [\(Table 4-1\)](#page-27-0).

Figure A-7. Dilution versus downstream distance for Case G [\(Table 4-1\)](#page-27-0).

Figure A-8. Dilution versus downstream distance for Case H [\(Table 4-1\)](#page-27-0).